

A critical perspective on interpreting amplicon sequencing data in soil ecological research

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1 Abstract

2 Microbial community analysis via marker gene amplicon sequencing has become a routine method in the field of soil
3 research. In this perspective, we discuss technical challenges and limitations of amplicon sequencing studies in soil and
4 present statistical and experimental approaches that can help addressing the spatio-temporal complexity of soil and the
5 high diversity of organisms therein. We illustrate the impact of compositionality on the interpretation of relative abundance
6 data and discuss effects of sample replication on the statistical power in soil community analysis. Additionally, we argue
7 for the need of increased study reproducibility and data availability, as well as complementary techniques for generating
8 deeper ecological insights into microbial roles and our understanding thereof in soil ecosystems. At this stage, we call upon
9 researchers and specialized soil journals to consider the current state of data analysis, interpretation and availability to
10 improve the rigor of future studies.

11

12 Highlights

- 13 ● Soil complexity necessitates careful interpretation of sequencing data
- 14 ● Studies often do not account for data compositionality, leading to misinterpretation
- 15 ● Functions should not be inferred from phylogeny as they are rarely conserved
- 16 ● We discuss complementary approaches that help to improve ecological insights
- 17 ● We call for journals and authors to improve study reproducibility and data availability

18 Keywords

19 amplicon sequencing, soil metabarcoding, soil microorganisms, soil microbial diversity, soil complexity,
20 compositional data

21 1. Introduction

22 Soil is one of the most biologically diverse and heterogeneous ecosystems, presenting unique challenges
23 to scientists in the fields of soil and microbial ecology (Bickel and Or, 2020). The critical role of mi-
24 croorganisms as drivers of biogeochemical processes is well-documented, and a major goal of soil ecology
25 remains to decipher the link between the diversity of soil microbial communities, and their function in the
26 environment (Hinsinger et al., 2009; Manzoni et al., 2012). Historically, studies of microbial communities
27 revealed rather a narrow perspective of the diversity by targeting mainly cultivable bacteria, taxa of high
28 abundance, or microorganisms grouped according to morphological or physiological properties (Staley
29 and Konopka, 1985; Steen et al., 2019; Åsa Frostegård et al., 2011). The introduction of next-generation
30 sequencing technologies such as amplicon sequencing has revolutionized our understanding of micro-
31 bial diversity by enabling the investigation of community composition at a much greater phylogenetic
32 resolution than ever before.

33 Amplicon sequencing (also termed metabarcoding) is based on PCR-amplification of variable regions
34 of DNA within conserved phylogenetic or functional marker genes (Gołębiewski and Tretyn, 2019; Se-
35 menov, 2021) - see also supplementary Table S1 for examples. The accessibility of established assays, the
36 affordability, as well as the availability of free analysis software packages have facilitated the broad use of
37 amplicon sequencing for characterization of the microbiological diversity in environmental samples (Ca-
38 poraso et al., 2012). In the field of soil science, its application has accelerated in the last decade as
39 evidenced by the growing number of studies published in specialized soil journals (Fig. 1). The majority
40 of these manuscripts report the analysis of soil community composition and diversity based on phyloge-
41 netic marker genes such as the 16S rRNA gene for bacteria and archaea as well as internal transcribed
42 spacer (ITS) regions for fungi. In addition, functional genes can be targeted to obtain information on
43 the organism that may contribute to a specific environmental process (Angel et al., 2018; Séneca et al.,
44 2020; Aigle et al., 2020).

45 Such work has enabled researchers to successfully investigate the composition and dynamics of soil
46 microbial communities. Our understanding of microbial diversity has increased dramatically and the
47 activity of microbial communities has now been widely recognized as central in the field of soil science
48 where research questions were historically often tackled from the perspective of individual disciplines such
49 as chemistry, physics, and biology (Baveye et al., 2018). As evident by the high number of studies being
50 published in recent years, it is safe to say that microbial community analysis via marker-gene sequencing
51 has become a standard tool in soil research. At this stage, it is necessary to discuss potentials, challenges,
52 and pitfalls of the technique applied by soil scientists.

53 In this perspective, we aim to describe the unique challenges of studying microbial communities in
54 soil ecosystems, and to address common misconceptions in the analysis and interpretation of amplicon
55 sequencing data. Patterns often arise in community data, but the interpretation of these patterns in a
56 soil context remains challenging and limited due to the poor link between the sequenced marker gene

57 regions and microbial functions, as well as the compositional nature of the data itself (Blanchet et
 58 al., 2020). We provide suggestions for designing sequencing experiments and analyzing data to gain
 59 improved insights into microbial community structure and dynamics within the context of the complex
 60 soil environment. Amplicon sequencing, when used as part of a well-designed experiment, represents
 61 an informative approach for investigating microbial community structure and correlations between taxa
 62 and environmental parameters, as well as for developing new hypotheses regarding microbial community
 dynamics.

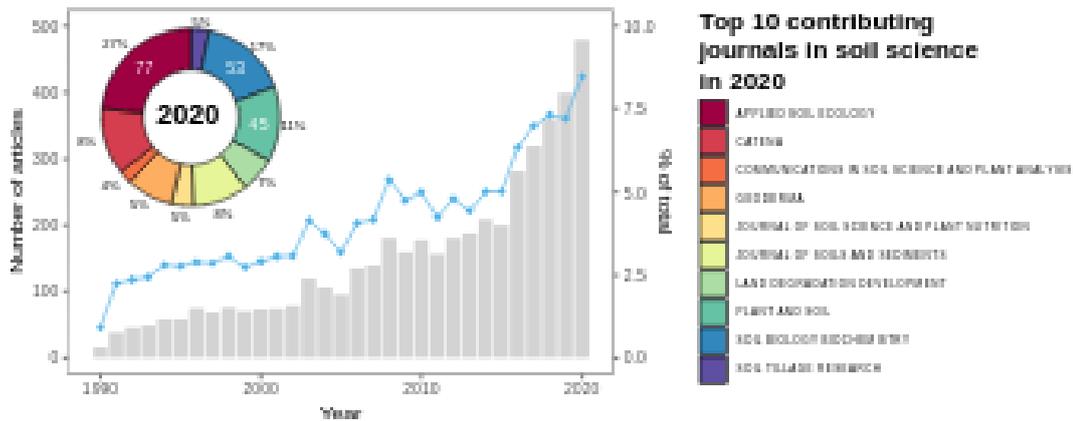


Figure 1: Increase in the number of articles using amplicon sequencing in soil microbiome research published in soil science journals (as defined in Web of Science, www.webofknowledge.com). Bars represent the total number of articles using amplicon sequencing, whereas the blue points and line represent their percentage of the total number of articles published in those journals per year from 1990 to 2020. The pie chart represents the number of articles in the top ten contributing soil science journals in 2020 (as total number of articles). Numbers inside the chart represent the number of articles using amplicon sequencing (only reported for the top three journals), while the numbers outside the chart represent the percentage of the total number of articles for each journal. See Supplementary file for a more detailed description of methods and the complete list of journals (Table S2).

63

64 2. Technical considerations in a heterogeneous and diverse habitat

65 The diversity of microorganisms in soil has been well-documented as a major challenge in studying
 66 soil microbial communities (Gans, 2005; Fierer and Jackson, 2006). A single gram of soil is estimated
 67 to contain 10^8 - 10^9 cells (Bloem et al., 1995; Nunan et al., 2001) and tens of thousands of microbial
 68 taxa (Roesch et al., 2007). Additionally, compared to host-associated microbiomes (e.g., gut, skin, or
 69 plant root microbiome), free-living bacteria exhibit higher levels of diversity. In a recent comparison
 70 of alpha-, beta- and gamma-diversity from samples collected as part of the Earth Microbiome Project

71 (EMP), soils were determined to have the highest alpha-diversity across all environments (Walters and
72 Martiny, 2020). In terms of beta- and gamma-diversity, soil came in second only to sediment samples.
73 Fewer studies have investigated the diversity and global distribution of fungi (Tedersoo et al., 2014;
74 Větrovský et al., 2019). These studies indicate that more heterogeneous environments, such as soils and
75 sediments, may contain more diverse fungal communities than more homogeneous habitats (e.g., marine,
76 freshwater, air, biofilms) (Fierer and Lennon, 2011; Walters and Martiny, 2020; Torsvik, 2002).

77 In addition to high biological diversity, researchers interested in the microbial composition of soils are
78 confronted with technical challenges throughout the sample processing workflow. The general workflow
79 of amplicon sequencing includes: 1) planning and implementation of the experimental design, 2) nucleic
80 acid extraction (including quality control) 3) primer choice, PCR amplification, sequencing, 4) processing
81 and analysis of sequence data, and 5) data interpretation (Fig. 2). At each of these steps, a subset of
82 the sample is selected and information can be lost as a result of the techniques applied (i.e., nucleic acid
83 extraction method, primer selection, statistical approaches), with consequences for data interpretation
84 in the context of ecological questions (Morton et al., 2019; McLaren et al., 2019). As with any scientific
85 experiment, the specific hypotheses to be addressed should determine the experimental design. Besides
86 this, in experiments involving amplicon sequencing, one must consider the appropriate spatial scale (i.e.,
87 aggregate/microscale, centimetre scale, meter scale) and the frequency of sampling in order to address
88 specific questions regarding community dynamics. While the sample that is sequenced represents the
89 specific moment in time when it was frozen or extracted, the presence of exogenous or relic DNA in
90 soil samples has the potential to influence community composition and downstream data interpretation
91 ((Lennon et al., 2018; Carini et al., 2016); discussed in section 5). Additionally, sample replication
92 remains a critical concern in soil studies, particularly when it comes to statistical inference and/or
93 construction of co-occurrence networks (discussed in sections 5 and 6).

94 The physicochemical properties of soils make nucleic acid extraction from this matrix particularly chal-
95 lenging. Numerous extraction protocols and kits have been developed to circumvent challenges with
96 DNA extraction from soil, however, each method introduces distinct bias on the subset of the microbial
97 community retrieved (Terrat et al., 2011; Zielińska et al., 2017; Dopheide et al., 2018). The presence
98 of inhibitors, such as humic substances, is common in soil and can reduce the quality and purity of
99 nucleic acids in the extracted samples and decrease the efficiency of reverse transcription and/or PCR
100 reactions (Schrader et al., 2012). In addition to the nucleic acid extraction method of choice (chemical
101 or physical lysis, DNA and/or RNA extraction), primer selection dictates the organisms or functions
102 targeted by the approach (phylogenetic or functional marker; see Table S1). Finally, due to the diversity
103 and heterogeneity of soil samples the resulting data is often sparse, containing numerous taxa with low
104 abundance and prevalence which may be dealt with through filtering thresholds or statistical approaches
105 (see section 3). The loss of information at each step of the process - from sampling to analysis - must
106 be carefully considered in light of amplicon sequencing data interpretation. Keeping all these factors in
107 mind, the application of sequencing technologies to soil has provided invaluable information regarding
108 the structure and critical nature of understanding microbial communities.

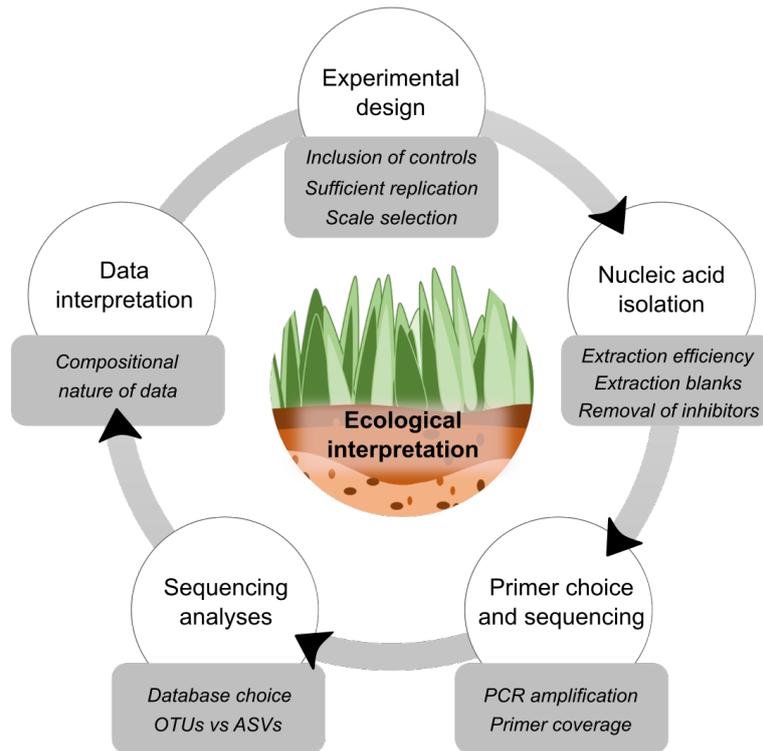


Figure 2: The main steps of an amplicon sequencing analysis workflow. Italicized items represent critical points that may strongly influence the robustness and direction of the results.

109 3. Challenges in amplicon sequence data analysis

110 3.1. Primer selection dictates phylogenetic coverage

111 As choice of primers can influence the taxa observed in an amplicon sequencing dataset, it is of utmost
 112 importance to take care with regard to primer selection and the interpretation of resulting data as to
 113 the community changes/impacts of treatments. Given the high diversity of soil communities (of which
 114 the understanding is constantly growing due to massive sequencing efforts), no primer pair will cover
 115 the complete phylogenetic breadth on a high rank such as domain (e.g., bacteria, archaea, fungi). As a
 116 consequence, part of the communities will always be missing. This is inevitable but it is of particular
 117 concern when studies attribute soil functions to taxa found to be “rare” in their amplicon sequencing
 118 data due to low coverage of that group (Chen et al., 2020). Nevertheless, evaluated and recommended
 119 primer pairs are available including the updated versions of 515F-806R primers for surveys of archaea
 120 and bacteria (e.g., <https://earthmicrobiome.org/>). Primer selection is even more challenging for
 121 studies of eukaryotes, owing to hypervariable sequence lengths and multiple gene copy numbers (due
 122 to multiple operons and/or polykaryosis) that contribute to biased amplification of some phylogenetic
 123 groups during PCR. This bias may for example lead to the under-estimation of some fungal groups,
 124 having a downstream effect on diversity estimates (Baldrian et al., 2021). Arbuscular mycorrhizal fungi
 125 for instance are largely overlooked by commonly used ITS primers which could lead investigators to infer
 126 that arbuscular mycorrhizal fungi are rare (George et al., 2019). A promising alternative to ITS-targeted
 127 short-read sequencing is the use of long-read sequencing (e.g., PacBio) which enables the investigation
 128 of most fungi (including *Glomeromycota*) and other soil eukaryotes through covering both the full ITS

129 region and part of the small subunit of the rRNA gene (Tedersoo and Anslan, 2019; Tedersoo et al.,
130 2020). We refer readers to in-depth reviews that further discuss challenges regarding amplicon sequencing
131 of fungi specifically, including discussion of primer selection and coverage (Nilsson et al., 2019; Baldrian
132 et al., 2021).

133 The choice of primers has substantial impacts on estimates of diversity in community studies. As a
134 consequence, we urge researchers to use tools such as TestPrime ([https://www.arb-silva.de/search/
135 testprime/](https://www.arb-silva.de/search/testprime/)) to evaluate the current status of the coverage of their target microbial groups of interest
136 before sequencing and to discuss this aspect in their publications. We also recommend that reviewers
137 critically assess the coverage of the target group of organisms used in a study to improve future evaluation
138 of sequencing-based research in soil ecology.

139 3.2. Compositionality necessitates careful data processing

140 One of the first steps in the analysis of amplicon sequencing data is the removal of potential sequencing
141 errors. Doing so eliminates sequencing artefacts that may falsely boost diversity levels (Edgar et al.,
142 2011; Haas et al., 2011). The use of amplicon sequencing variants (ASVs), instead of operational taxo-
143 nomic units (OTUs) helps to overcome this issue by assigning a greater probability of a true biological
144 sequence being more abundant than an error-containing sequence (Callahan et al., 2017). To that end,
145 bioinformatic tools such as DADA2 (Callahan et al., 2016) and Deblur (Amir et al., 2017) attempt to
146 use sequencing error profiles to resolve amplicon sequencing data into ASVs. An ASV is more likely to
147 have an intrinsic biological meaning (i.e., being a true DNA sequence), as opposed to an OTU which can
148 either be a representation of the most abundant biological sequence or a consensus sequence (of which
149 the latter may not exist in reality). In addition, ASVs facilitate the merging of datasets, particularly
150 when the same sequencing primer pairs are used.

151 Another relevant step when analyzing sequencing data is to account for the different sequencing efforts
152 across samples (i.e., sequencing depth) that can result in a substantially different number of recovered
153 reads even among replicates. Ways to tackle this include total library size normalization and rarefac-
154 tion, with both remaining debated to date (McMurdie and Holmes, 2014; Weiss et al., 2017). Bioin-
155 formatic tools such as DeSeq2 and EdgeR provide ways to normalize count tables (Love et al., 2014;
156 Robinson and Oshlack, 2010). Both methods are applied on raw or low-abundance filtered count tables,
157 and have performed well in both real as well as simulated datasets and outperform rarefaction-based
158 approaches (McMurdie and Holmes, 2014). Other alternatives that account for the compositional as-
159 pect of sequencing data include centered log-ratio (CLR), isometric log-ratio (ILR) or additive log-ratio
160 (ALR) ratios transformations on a count data matrix with adequate replacements of zeros (Aitchison,
161 1984; Egozcue, 2003).

162 Following data normalization, traditional workflows include the generation of distance matrices for or-
163 dination, clustering, and variance partitioning analyses. Commonly used distance metrics include Bray-
164 Curtis, Jaccard and Unifrac (weighted and unweighted). These metrics are often used although they do
165 not take into account the compositional nature of sequencing data. The Aitchison distance - defined as
166 the Euclidian distance on top of a centered log-ratio transformed count matrix - is a viable composi-
167 tional alternative (Aitchison, 1984) that allows performing ordinations (e.g., PCA biplots). Additionally,
168 the “Phlir” transformation metric has been introduced as a compositional alternative to the weighted
169 Unifrac that carries phylogenetic information (Silverman et al., 2017). Most of the above mentioned
170 compositional options are implemented in R packages and include publicly available tutorials. In light

171 of the challenges related to normalization and analysis of compositional data, we recommend a critical
172 evaluation of available data analysis tools to best address the nature of each experimental setup (see also
173 section 6).

174 Another aspect that prevents data analyses from being fully quantitative is the potential of multiple copies
175 of marker genes present per organism, which may also vary across taxa. For example, the 16S rRNA gene
176 copy number per bacterial cell can vary between 1 and 18 and can additionally show variation within
177 different strains of the same species (Stoddard et al., 2014; Coenye and Vandamme, 2003; Lavrinienko et
178 al., 2021). Therefore, relying solely on the number and diversity of markers such a 16S rRNA genes can
179 lead to inaccurate estimates of the relative abundance and diversity of microbial communities. Several
180 computational tools can correct amplicon datasets for the number of 16S rRNA gene copies based on
181 existing genome information (e.g., PICRUSt2 (Douglas et al., 2020) and CopyRighter (Angly et al.,
182 2014)). However, correcting for 16S rRNA gene copy numbers in sequencing surveys remains challenging,
183 particularly for soil, as the gene copy numbers are only known for a subset of the soil microbes (Louca et
184 al., 2018; Nunan et al., 2020). This challenge becomes even more problematic for marker genes of fungi
185 and other eukaryotes, such as protists, as the copy number here can vary drastically between taxa (Gong
186 et al., 2013; Gong and Marchetti, 2019). Other housekeeping genes, which occur only once in a genome,
187 have been proposed as universal phylogenetic marker genes (such as *recA* (Eisen, 1995)), but their use
188 remains limited due to lower phylogenetic resolution and limited availability in databases.

189 3.3. Insufficient data availability contributes to a lack of reproducibility

190 Reproducibility and reusability of research results are predicated on sharing data and analysis scripts, a
191 topic of growing relevance in light of increasing amounts of sequencing data obtained from soils around
192 the globe and with the increasing complexity of analyses. Proper data sharing practices allow researchers
193 to re-analyze specific aspects of published datasets, and/or investigate patterns in soil communities
194 across datasets in the form of meta-analyses. A prerequisite to ensure data storage and availability
195 in a usable format is that authors are required to do so by respective journals. In order to assess the
196 current state of data deposition in the field, we searched the author guidelines of the 10 specialized soil
197 journals (see Fig. 1 for reference). Out of the 10 journals, many “encourage their authors to make data
198 available” while only 2 journals specifically require sequencing data to be deposited in public repositories
199 such as GenBank before a manuscript is accepted for publication. Even if authors feel encouraged to
200 comply, storage of their data in a repository does not always facilitate reproducibility of the reported
201 research. Deposited datasets often contain only raw results from whole sequencing runs, and provide
202 little meaningful information on the individual amplicons and on the corresponding metadata. As a
203 consequence, it may be difficult to reconstitute the exact datasets used for the reported statistics and
204 illustrations from such data. This requires that the applied quality filters and processing steps (see
205 section 3.1), as well as the versions of applied software packages, be precisely reported.

206 Consequently, we call on all specialized soil journals that accept and publish sequencing data to (i) provide
207 community standards for reproducible data analysis in their data policy statements and (ii) require the
208 submission of sequencing data, ASV/OTU tables, together with sample metadata, to open repositories
209 (such as GenBank, Dryad, or FigShare) and (iii) require that analysis scripts be made available on
210 open hosting services (such as GitHub) or accompany the publication as a supplement. These steps will
211 greatly facilitate reproducibility, open science, and meta-analyses.

212 4. Addressing and interpreting compositional sequencing data

213 4.1. Interpreting relative abundance data

214 The compositionality of amplicon sequencing data presents challenges to the interpretation of changes
 215 in microbial community structure. The amount of sequence data obtained through high-throughput
 216 sequencing is a fixed value, resulting in a random sampling of sequences from a sample that cannot
 217 be directly linked to absolute abundance based on sequences alone (Gloor et al., 2017). Numerous
 218 studies have revealed shifts in microbial community composition across treatments including gradients
 219 of temperature, pH, and salinity, as well as seasonal or temporal parameters. This practice is robust on
 220 a community level when broad-scale changes in taxa are of interest (e.g., phylum level), and has resulted
 221 in similar ecological conclusions as data generated with more quantitative approaches (Piwosz et al.,
 222 2020). However, at higher taxonomic resolution (e.g., genus level), quantitative inferences from relative
 223 abundance sequencing data become more challenging. Due to the nature of sequencing, a change in the
 224 relative abundance of one species is always reflected in a corresponding change in one or more other
 225 species. We depict such challenges in interpretation in the following thought experiment (Fig. 3).

226 Amplicon sequencing data obtained from the same soil sample at two different time points (t_1 , t_2)
 227 consists of two species (A, B). The relative abundance observed for species A and B is 0.55 and 0.45 at
 228 time point 1 (t_1), and 0.8 and 0.2 at time point 2 (t_2), respectively (Fig. 3). From t_1 to t_2 , species B
 229 decreases in relative abundance coupled to an increase in the relative abundance of species A. The bars
 230 below (t_2a - t_2e) illustrate five examples of changes in absolute abundance in t_2 that could underlie the
 231 patterns observed in relative abundance data. The initial time point (t_1) is also shown for comparison.

232 The first case represents a situation where the absolute abundance matches the relative abundance
 233 observations. There are no changes in total biomass from t_1 to t_2 and species A increases, whereas
 234 species B decreases (Fig. 3, t_2a). The second case depicts an increase in overall biomass between t_1
 235 and t_2 caused by an absolute increase in species A and no absolute changes in species B (Fig. 3, t_2b).
 236 The third case represents an opposite scenario where the decreases in total biomass between t_1 to t_2 is
 237 caused by a decrease in species B and no changes in species A (Fig. 3, t_2c). The fourth case represents
 238 a situation where there is a general increase in biomass from t_1 to t_2 prompted by increases in absolute
 239 abundances of both species A and B (Fig. 3, t_2d), while the fifth case represents an opposite scenario
 240 (Fig. 3, t_2e). For some of these examples, observed changes in relative abundance may accurately reflect
 241 true biological changes (t_2a , t_2d and t_2e), whereas interpretation of the community shifts that underlie
 242 observed patterns remains more difficult for the other scenarios (t_2b and t_2c). Without information on
 243 absolute abundances, there is still room for ambiguous interpretations solely based on relative abundance
 244 plots (see section 4.2). This theoretical exercise shows, that even for a community of only two member
 245 species, there are five potential scenarios of changes in the absolute abundance that could cause the
 246 observed shift in relative abundance. Given that soil communities usually harbour thousands of species,
 247 the degree of complexity increases dramatically.

248 4.2. Experimental approaches to address compositionality

249 The challenge of interpreting relative abundance data as illustrated in Figure 3 indicates the advantages
 250 of adding quantitative information to current amplicon sequencing approaches. Knowledge on absolute
 251 values (e.g., total microbial biomass) can help to make more robust inferences about the nature of

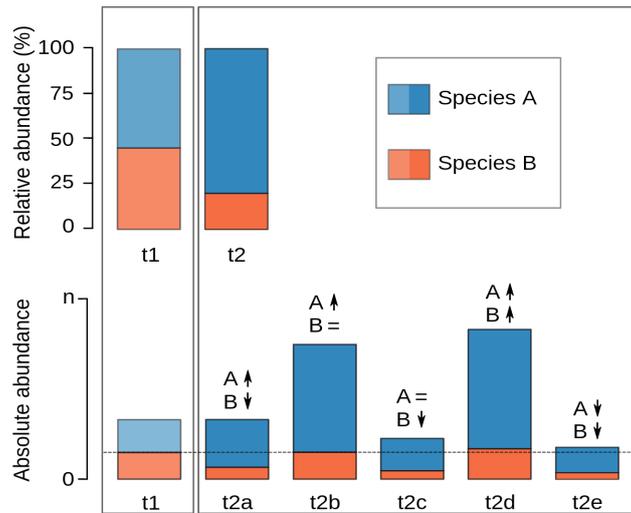


Figure 3: Relationship between the relative abundance of species as observed via amplicon sequencing and their absolute abundances. The upper panel shows the relative abundance of two species A (shades of blue) and B (shades of pink) at two time points in a theoretical experiment. From t1 to t2, a decrease in relative abundance of species B is observed, coupled to an increase in relative abundance of species A. The relative abundance pattern observed at t2 could have been caused by five changes in the biomass and absolute abundance of the microbial community as shown in the lower panel. Time point t1 is shown for comparison.

252 observed shifts in microbial community structure (Fig. 3, t2d and t2e; (Barlow et al., 2020; Wang et al.,
 253 2021). In the following, we discuss some approaches ranging from molecular techniques to classic soil
 254 microbiology that could help improve our interpretation of amplicon sequencing data.

255 4.2.1. Quantitative PCR approaches

256 One relatively affordable and well-established quantitative method is quantitative real-time PCR (qPCR).
 257 qPCR enables to assess copy numbers of a marker gene which may be multiplied by the relative abundance
 258 data of the same sample obtained by amplicon sequencing. This approach benefits strongly from using
 259 the same primers in both qPCR and sequencing to reduce bias stemming from PCR (see section 2) and
 260 from correcting for the copy numbers of said marker gene in the genome of target organisms.

261 A relatively novel alternative to traditional qPCR is digital PCR (dPCR) which requires no exter-
 262 nal standard for quantification, offers higher precision, and is relatively unaffected by the presence of
 263 PCR inhibitors. This represents a tremendous advantage when working with nucleic acid extracts from
 264 soil (Dong et al., 2015). However, like standard qPCR, the efficiency of this method is affected by the
 265 degeneracy of the primers, which means particular care must be taken during primer design (see section
 266 3.1). In addition, both dPCR and qPCR are limited in terms of absolute quantification of the fungal ITS
 267 gene due to the hypervariable target region and its variable-length (Nilsson et al., 2019).

268 A major advantage of both quantitative PCR approaches is the possibility of using the same DNA
 269 extracts as for the community profiling without additional sample processing that would be required

270 for other methods (see sections 4.2.2- 4.2.4). Consequently, quantitative PCR approaches have been
 271 used successfully to address the compositionality of sequencing data and can aid in the interpretation of
 272 microbial community data in soil (e.g., (Tkacz et al., 2018; Zemb et al., 2020; Vandeputte et al., 2017;
 273 Kleyer et al., 2017)) .

274 4.2.2. Spike-ins

275 Introducing an internal standard (also called a spike-in) can be a useful tool toward achieving more
 276 quantitative amplicon data analyses. Spike-ins can be introduced in the form of microbial cells (Stämmeler
 277 et al., 2016) or as selected DNA sequences (Tkacz et al., 2018; Hardwick et al., 2018; Wang et al., 2021).
 278 The spike should be uniquely detectable as a non-member of the existing microbial community, and should
 279 not be introduced in concentrations that would shift the sequencing effort towards it. Additionally,
 280 the timing of the addition will determine the type of information retrieved. While adding the spike
 281 after extraction can provide good estimates of amplification and/or sequencing biases, it does not take
 282 extraction efficiency into account (Hardwick et al., 2018; Stämmeler et al., 2016). A recent amplicon
 283 sequencing study applied a synthetic DNA spike of known concentration to faecal samples prior to
 284 extraction. They combined this with qPCR quantification to calculate the number of gene copies after
 285 accounting for the extraction yield. The ratio of each OTU against the initial concentration of 16S rRNA
 286 genes was used to calculate more accurate abundance levels of each OTU after taking extraction efficiency
 287 into account (Zemb et al., 2020). If performed in a comparable manner, spike-ins represent a promising
 288 tool to determine abundances of taxa more quantitatively via sequencing in future soil studies.

289 4.2.3. Direct cell counts

290 Another approach towards absolute abundance data from soil communities are direct cell counts obtained
 291 through fluorescence microscopy (Bloem et al., 1995) or fluorescence-activated cell counting (Khalili et
 292 al., 2019; Frossard et al., 2016) of cells liberated from soil particle surfaces (Riis et al., 1998; Lentendu
 293 et al., 2013). Total counts help to assess the absolute abundance of microbial cells that fall within a
 294 certain range of parameters such as cell size and morphology. Cell counting approaches remain more
 295 straightforward for single-cell archaea and bacteria than for filamentous bacteria, fungi or other soil
 296 eukaryotes. The success of cell counting can be negatively affected by soil autofluorescence (low signal-
 297 to-noise ratio), partial separation of microbial cells from soil particles, or masking the detection of cells
 298 by overlaying soil particles. Nevertheless, assessing the number of cells in samples also subjected to
 299 sequencing may help to estimate changes in absolute abundance and to better interpret sequencing data
 300 (Fig. 3).

301 In addition, the observation and enumeration of target species of interest through marker-based ap-
 302 proaches (e.g., FISH: fluorescence *in situ* hybridization) enables the quantification of absolute abundances
 303 of those species identified through sequencing. This practice not only allows soil ecologists to verify if the
 304 change observed in relative abundance indeed translates to shifts in the community by counting taxa of
 305 interest on filters (Piwosz et al., 2020), but also expands the interpretation of sequencing data to localize
 306 and visualize species of interest *in situ* (e.g., on roots (Martin et al., 2020)) and to elucidate ecological
 307 implications behind changing abundances of target species in soil samples. Applications of FISH in con-
 308 junction with amplicon sequencing to soil samples are surprisingly rare albeit such targeted localization
 309 and enumeration is a powerful tool to understand the dynamics of certain phylogenetic groups in soil on
 310 a quantitative basis.

311 4.2.4. Combining classical soil biogeochemical methods with amplicon sequencing

312 Traditional soil biogeochemical approaches enable the quantification of total microbial biomass in soil,
313 including methods such as chloroform fumigation extraction (CFE)(Brookes et al., 1985), phospholipid
314 fatty acid (PLFA) profiling (Frostegård et al., 1991; Åsa Frostegård et al., 2011; Buyer and Sasser,
315 2012) and ergosterol measurements (Joergensen and Wichern, 2008; Montgomery et al., 2000). In con-
316 trast to PCR-based methods, they assess the concentration of chemical microbial biomarkers in soil
317 directly, thereby avoiding biases introduced by amplification of the target molecules. For example,
318 such quantitative information regarding an increase or decrease in total microbial biomass between
319 treatments would complement corresponding shifts in relative abundance data as observed via amplicon
320 sequencing (Fig. 3). In addition to assessing total microbial biomass, PLFA measurements can also
321 generate abundance information for microorganisms at a coarse phylogenetic resolution. The ability to
322 obtain abundance profiles for bacteria, fungi, as well as distinguishing between gram-positive, gram-
323 negative, and *Actinobacteria*, could be used as a “benchmark” for interpreting relative abundance data
324 for more specific subsets of an amplicon dataset (i.e., (Drigo et al., 2010)). A combined interpretation of
325 datasets from biochemical and molecular methods with fundamentally different measurement principles,
326 however, may not always be as straightforward as the combination of amplicon sequencing data with
327 quantitative PCR (see section 4.2.1).

328 Overall, we suggest that adding any quantitative measurement of microbial abundance such as quantita-
329 tive PCR, cell counting, CFE, or PLFA will benefit and guide the interpretation of amplicon sequencing
330 data. The use of more quantitative tools will provide a more robust foundation to reduce misinterpreta-
331 tion of compositional sequencing data by providing a link between total microbial biomass and changes
332 in the relative abundance of microbial groups.

333 5. Linking sequences to ecological context

334 5.1. Soil spatial complexity occurs on micro- and macro- scales

335 Investigating microbial community composition in soils presents unique challenges. Compared to well-
336 mixed ecosystems, microbial life (i.e., growth, activity, dormancy, and turnover) in the soil is strongly lim-
337 ited by the complex network of pores, as well as gas transport and diffusion in the aqueous phase (Bickel
338 and Or, 2020; Young, 2004; Vos et al., 2013). Soil microarchitecture is a key factor that influences the
339 potential for microorganisms to interact with each other (Wilpiseski et al., 2019). In practice, how-
340 ever, the analysis of soil microbial communities through amplicon sequencing does not account for soil
341 microarchitecture. Researchers commonly use bulk homogenization approaches to extract nucleic acids
342 from 250 - 500 mg of fresh soil which naturally obscures the physical structure and spatial arrangements
343 of microbial cells in this soil sample. From the microbial perspective, nucleic acid extraction represents a
344 macroscopic measurement of the “whole” microbial community. This practice does not negatively affect
345 soil microbiome analyses unless interactions among microbial taxa are inferred (e.g., via network analysis,
346 see section 5.4).

347 The spatial heterogeneity of soil and the microbial communities therein does not only persist on the
348 microscale, but certainly also on a centimeter, meter, field, or ecosystem scale (Becker et al., 2006; Wolfe
349 et al., 2006; Franklin and Mills, 2003). Sampling “the same soil” a few meters apart or at different depths

350 in the soil profile might result in individual samples with varying biogeochemical properties such as pH,
 351 water saturation, soil texture, and also plant root distribution (Zhang and Hartemink, 2021). Choosing
 352 a sufficient number of replicates to assess sample or plot variability while balancing the cost-to-gain ratio
 353 is certainly an important measure to address soil heterogeneity (see section 6). Thus, it is critical to
 354 carefully evaluate the representativeness of technical and biological replicates. A recent study showed
 355 distinct and consistent differences in bacterial and fungal communities between individual replicate soil
 356 samples throughout a season even though 10-15 cores were randomly sampled in individual subplots and
 357 pooled (Carini et al., 2020). Another study showed that chemical soil properties, as well as microbial
 358 biomass and communities, exhibited high levels of spatial variation across 49 samples in a 6 × 6 m
 359 forest plot (Štursová et al., 2016). The pooling of samples, individual extractions of DNA/RNA and/or
 360 amplification reactions made from a single DNA template can certainly dampen confounding effects of
 361 community heterogeneity. Nevertheless, existing intraplot variability and representativeness of samples,
 362 as well as the appropriateness of sampling strategies to correctly address them, must be critically assessed
 363 in any study on soil microbiomes. Otherwise, drawing of generalized macroecological conclusions from
 364 soil samples taken and pooled across large distances may yield speculative information at best (Zhang
 365 et al., 2020; Dini-Andreote et al., 2020).

366 5.2. Temporal scales to consider when analyzing microbial dynamics

367 When designing an experiment, one must not only consider the spatial scales at which microorganisms
 368 live and interact but as well the temporal scale, i.e., the frequency at which sampling should occur to
 369 capture temporal dynamics. Amplicon sequencing represents a snapshot of microbial prevalence at a
 370 given moment. Given that microbial community turnover among different soils is may range from weeks
 371 to years (e.g., (Spohn et al., 2016)), it is difficult to assess the best temporal sampling strategy *a priori*. If
 372 for example effects of root exudation on soil microbial community dynamics are of interest, it is important
 373 to consider the different temporal scales of the processes to be correlated. Root exudation varies with
 374 plant development stage and shows diurnal patterns (Oburger et al., 2014), whereas community changes
 375 on a DNA level may not be detectable on such a short temporal scale (in contrast to RNA, see below).
 376 Any pattern of a single sampling time point would rather represent a legacy community that established
 377 around plant roots instead of the current state of a community that can be linked to root exudation
 378 (composition, rate) measured at the same time point.

379 Another soil parameter that might mask the detection of community shifts is intrinsically linked with
 380 microbial turnover: relic or environmental/exogenous DNA. Relic DNA is extracellular DNA from non-
 381 viable cells that has leaked into the environment and that is thought to persist in soils for months to
 382 years (Levy-Booth et al., 2007; Carini et al., 2016). Relic DNA has been estimated to comprise approxi-
 383 mately 40% of the amplifiable soil DNA pool and has been successfully removed from soil samples via the
 384 application of DNAses or propidium monoazide (Lennon et al., 2018; Carini et al., 2020; Carini et al.,
 385 2016). The latter study found greater differences in soil communities across several time points where
 386 relic DNA was removed as compared to samples where relic DNA was still present. Consequently, the
 387 presence of relic DNA may complicate the interpretation of sequencing data by over- or under-estimating
 388 microbial diversity which may be of particular concern when temporal dynamics are key to the scientific
 389 question.

390 One possibility to address short temporal dynamics while eliminating bias of relic DNA is ribosomal RNA
 391 (rRNA) amplicon sequencing via complementary DNA (cDNA) synthesis. The lifetime of rRNA in soils

392 is relatively short and has been estimated to range from days to a few weeks depending on biogeochemical
393 parameters such as temperature, pH, and water saturation (Schostag et al., 2020; Blazewicz et al., 2013).
394 Thus, rRNA-targeted amplicon sequencing may increase the chances of capturing dynamics within soil
395 microbial communities over time and may be used to carefully assess the “active” fraction thereof (Vieira
396 et al., 2019) (see Table S2). Caution should still be taken when sequencing of nucleic acids at higher
397 frequencies, even if relic DNA has been removed or RNA is used. If community dynamics are to be
398 investigated in short time intervals (e.g., minutes to hours) we suggest combining amplicon sequencing
399 with methods for targeting the metabolically active cell fraction (as discussed in section 7).

400 5.3. Inferring function from phylogeny

401 Although some links exist between the environment and the community composition therein, amplicon
402 sequencing cannot be used to predict microbial function and roles in ecological processes (Fierer et al.,
403 2007; Fierer, 2017). Nevertheless, it can serve as a useful tool to survey microbial communities through
404 detection of a section of a single gene or gene region (Fig. 4). The consequence of targeting a subsection
405 of microbial genomes is that ecological insights that can be extracted from these data remain limited.
406 Function of taxa identified via amplicon sequencing cannot simply be inferred from the phylogeny of these
407 organisms, as complex evolutionary processes (e.g., horizontal gene transfer) play a key role in functional
408 trait distribution across the genomes of microorganisms (Menna and Hungria, 2011). Function may not
409 necessarily be conserved across phylogenetic levels, and therefore processes cannot be reliably predicted
410 and assigned to taxa using amplicon sequencing targeting phylogenetic markers such as 16S rRNA
411 genes (Nunan et al., 2020; Li et al., 2019). Consequently, we suggest to avoid inferring life strategies of
412 taxa via their classification into a phylum (e.g., equating *Proteobacteria* with fast-growing r-strategist)
413 and using such assumptions to explain processes in soils for surveys based on general markers such as 16S
414 rRNA genes (Jeevani et al., 2020) and ITS regions (Zhou et al., 2021).

415 Recent studies apply functional predictions using packages such as PICRUSt2 (Douglas et al., 2020) or
416 Tax4Fun (Aßhauer et al., 2015), which suggest that metagenomes (and therefore functional potential of
417 organisms) can be extrapolated from the sequenced amplicon using phylogenetic markers. In the case
418 of fungi, FUNGuild or FungalTraits have been developed, which parses OTUs/ASVs into functional
419 guilds based on similarity to existing reference sequences (Nguyen et al., 2016; Pölme et al., 2020). The
420 main limitation of these approaches lies in the fact that they are dependent on a single gene, and the
421 completeness of reference sequence databases, many of which remain incomplete due to bias in the types
422 of organisms for which we have references (section 3, (Choi et al., 2016)). However, these prediction-
423 based software packages can be used to generate valuable hypotheses for further investigation or an
424 additional line of evidence to support a finding. In such cases, we recommend to follow up by either
425 FISH-counting of the identified species, functional gene-targeted sequencing, or SIP experiments to learn
426 more about the species or community that is hypothesized to be responsible/involved in an ecosystem
427 process (further discussed in section 7).

428 5.4. Interpreting co-occurrence data and networks

429 Challenges associated with amplicon sequencing analysis and interpretation also complicate the use of co-
430 occurrence network analysis from soil samples. Generally, co-occurrence analysis generates networks with
431 biological species as nodes and edges representing associations between them. Network construction is
432 based on the detection of significant correlations between taxa, and can be used to investigate properties

433 of microbial communities including organismal co-existence (e.g., (Barberán et al., 2011)), identification
434 of keystone species (e.g., (Banerjee et al., 2018)) and the stability of community structure (e.g., (de
435 Vries et al., 2018; Shi et al., 2016)). There has been a recent upsurge in the number of studies including
436 the construction of association networks for soil microbial communities. However, many of these studies
437 have been criticized for their highly descriptive use of networks, that do not allow for an ecological
438 interpretation of detected patterns.

439 The difficulty in interpretation stems from inferring causal relationships between taxa based on corre-
440 lations, which is a long-standing topic of discussion in ecology (Blanchet et al., 2020; Barner et al.,
441 2018). Particularly for soil, it is important to keep in mind that the data contained in each environ-
442 mental sample is only a snapshot of complex spatio-temporal dynamics (see sections 5.1 and 5.2). As
443 interactions occur at the level of individual microorganisms, inferring interaction among microorganisms
444 in soil is facilitated if samples were taken on the microscale or aggregate scale, rather than on the bulk
445 or horizon scale (see Fig. 4). Independent from scale, any sequencing data from soil capture a noisy
446 signal which reflects several biological processes including: reproduction, death, dispersal, environmental
447 filtering, as well as intra- and inter-specific interactions. The heterogeneity (and resulting sparsity) of
448 amplicon datasets represents an additional confounding effect that may introduce spurious associations,
449 posing additional challenges unique to the study of soil ecosystems.

450 For microbiome data, the associations are most often assigned through the detection of significant cor-
451 relations between relative abundances, where spurious links can be detected if compositional data is not
452 appropriately handled (as explained Section 4). Several popular network construction tools, including
453 SparCC (log ratios) and SPIECEASI (clr), apply log ratios to address compositionality in the process
454 of network construction (Kurtz et al., 2015; Friedman and Alm, 2012). Another option is to convert
455 relative abundances into absolute values by using the total gene copy numbers obtained from qPCR
456 (see section 4). To improve this analysis we suggest a careful comparison of data with null models to
457 help interpret the results and eliminate some indirect associations between species (Connor et al., 2017).
458 Additionally, the use of complementary environmental measurement data can improve ecological insights
459 from networks (Goberna et al., 2019; Lima-Mendez et al., 2015). We recommend performing follow-up
460 experiments to further investigate potential interactions to explore inferences made through network
461 analysis. In summary, the field of network inference is rapidly evolving and alternatives are emerging
462 to address currently standing issues. Nevertheless, we still lack a definite framework that allows for a
463 straightforward interpretation of generated co-occurrence networks.

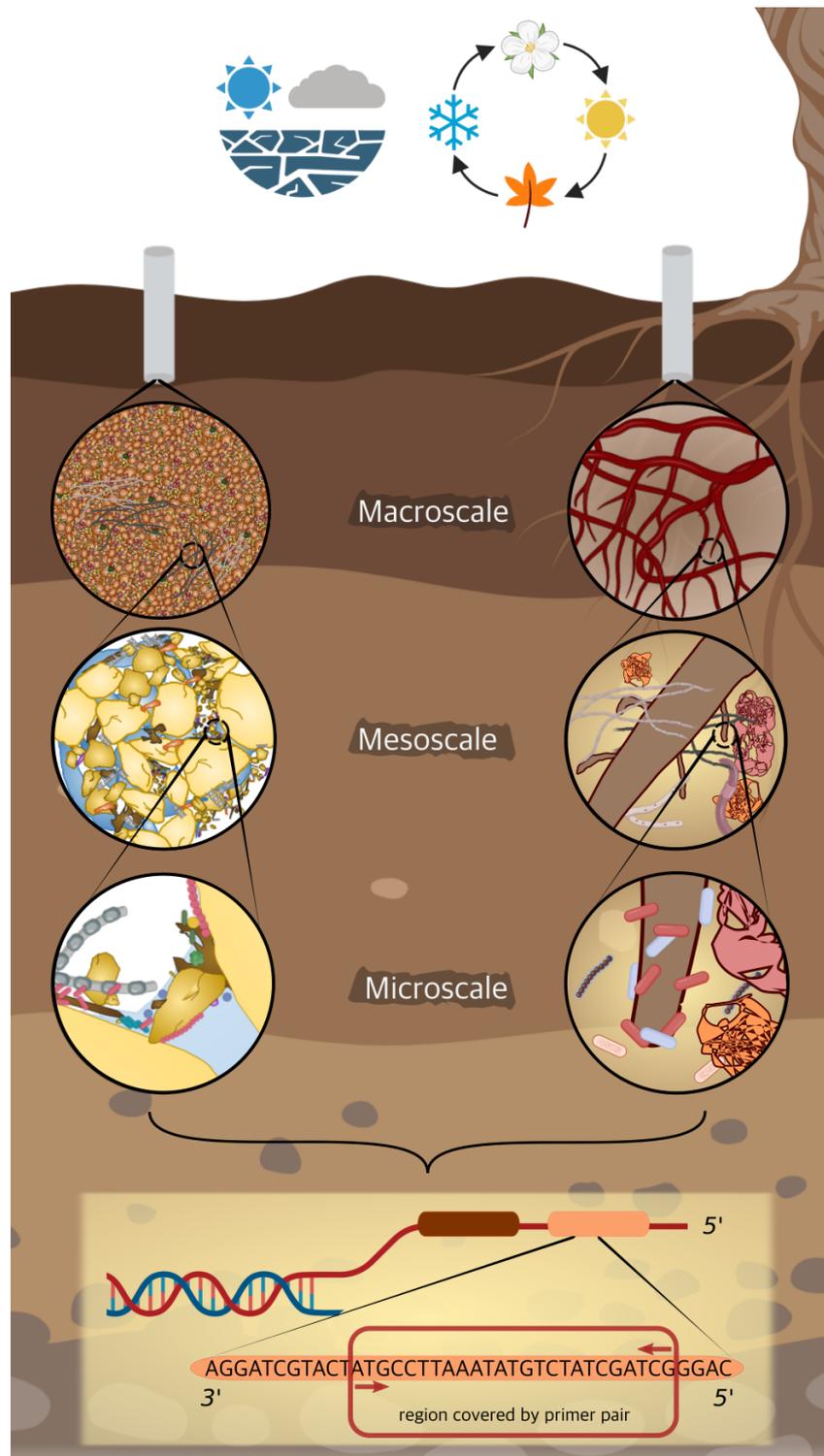


Figure 4: Schematic representation of the main spatio-temporal scales of soil ecosystems. Climate and seasonal patterns are depicted aboveground. The three main scales at which researchers investigate soil microbial communities are depicted as the macroscale, mesoscale, and microscale. Circle insets show the resolution at which microbial communities can be studied at each scale, emphasizing that careful experimental planning must be undertaken to capture community dynamics of interest. A partial single region of a selected marker gene that is captured by amplicon sequencing is depicted in the lower yellow box.

464 6. Addressing the overinterpretation of sequencing data

465 Amplicon sequencing data are well-suited for exploratory analysis and hypothesis generation in soil
466 research, but can also be applied for targeted hypothesis testing if appropriate complementary and
467 statistical methods are selected ((Gloor et al., 2017); sections 3 and 4). As amplicon datasets from soil
468 are characterized by compositionality, heterogeneity and sparsity, the use of standard statistical methods
469 (including Pearson correlations or *t-tests* on proportions) can lead to very high false-positive discovery
470 rates (up to 100% ; (Mandal et al., 2015; Morton et al., 2017)). Almost any soil microbiome data set will
471 show significant correlations as the data consist of thousands of individual variables. The possibility to
472 obtain significant results, therefore, may also lead to an abuse of the statistical significance (also referred
473 to “p hacking”). These effects are further compounded by spatio-temporal dynamics that contribute to
474 challenges in statistical inference from amplicon sequencing in soils (see section 5). Consequently, we ask
475 researchers to apply caution when inferring effects or associations solely based on statistical significance.
476 The recent discussion surrounding the abuse of p-values has resulted in alternatives and suggestions for
477 the use of more stringent p-values to reduce the false-positive discovery rate (Nuzzo, 2014; Amrhein et
478 al., 2019; Wasserstein et al., 2019; Benjamin et al., 2017). This would require an estimated dramatic
479 increase in sample size (up to 70%), which would be costly, but could also save money in the long run
480 that would have been spent on unsubstantiated research.

481 We explored the impact of sample replication on statistical power in soil microbiome analysis using a
482 published dataset on bacterial and fungal communities that features a range of soils representative of the
483 heterogeneity and biological diversity of soils (Zheng et al., 2019) (see supplementary methods) following
484 the approach described in (Kelly et al., 2015). We simulated OTU/ASV tables (see supplementary
485 information for description of data processing) and computed the dependency of statistical power of
486 permutational multivariate analysis of variance (PERMANOVA) on the effect size, by bootstrapping the
487 simulated matrices with varying replicate numbers (4, 5, 8 and 10 replicates; Fig. 5). We briefly
488 described the procedure used in the Supplementary information and address the reader to previous
489 publication (Kelly et al., 2015) for further details and how to implement the analysis with the package
490 ‘micropower’ available for R programming language.

491 Figure 5a shows the statistical power to detect significant differences with increasing effect size for
492 multiple groups (representing different sample sizes). This clearly shows that even a small increase in
493 the sample size increases the power to detect small differences. These results are similar to the findings
494 described in (Kelly et al., 2015) using the Human Microbiome Project (HMP) dataset with 16S rRNA
495 marker gene data sampled at multiple body sites. To better visualize these differences, we further
496 calculated the average statistical power for a range of effect sizes (ω^2) defined as ‘Low’ (0.001-0.04),
497 ‘Medium’ (0.04-0.08) and ‘High’ (0.08-0.12). Our analysis showed that the number of replicates hardly
498 affects the statistical power if there was a strong effect of treatment/site(Fig. 5b, “High”). However, if
499 the simulated treatment/site effect was lower, we found that an increase of the replicate number from 4
500 to 5 was sufficient to almost double the statistical power of small effect size (“Low”) and to achieve the
501 recommended power above 0.8 for medium effect sizes (Fig. 5b, “Low” and “Medium”). Consequently,
502 these effects were more pronounced when the number of replicates was doubled (4 to 8; Fig. 5b). Identical
503 effects were observed for the fungal data set (Fig. S1bc).

504 In practice, obtaining knowledge about the level of differences in soil microbial communities *a priori* is a
505 complicated undertaking. If preliminary sequencing data is available we encourage researchers to perform

506 such power analyses before experimental planning. Such considerations should also include the amount
 507 of technical replicates that will be pooled to alleviate the spatial heterogeneity of soils (see section 5).
 508 We refer to further literature on experimental planning and robust statistical analyses (e.g., (Coenen et al., 2020; Kelly et al., 2015; Johnson et al., 2014)).

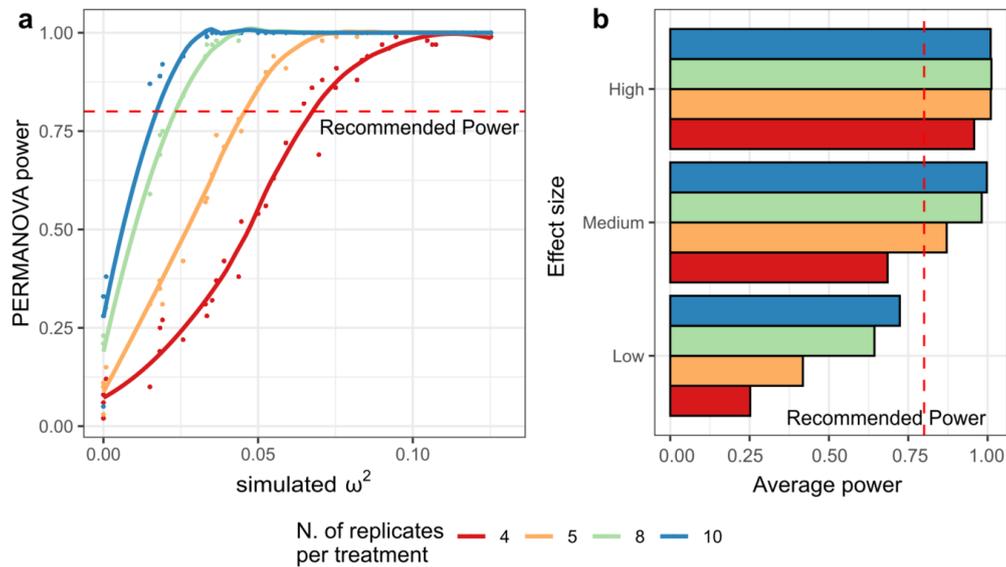


Figure 5: Graph showing: a) the calculated PERMANOVA power for a range of simulated effect (quantified by the adjusted coefficient of determination ω^2) and divided by number of replicates per treatment); b) the average PERMANOVA power of panel ‘a’, grouped by number of replicates per treatment and into three effect size ranges: Low (0.001-0.04), Medium (0.04-0.08) and high (0.08-0.12). PERMANOVA power was calculated as the proportion of bootstrap distance matrices for which PERMANOVA P -values are less than the pre-specified threshold for type I error (0.05)

509

510 7. Complementary approaches to amplicon sequencing that im- 511 prove ecological insights

512 As a consequence of the relative nature of amplicon sequencing data, the majority of such studies
 513 are descriptive. Marker-gene base surveys have certainly contributed to generate valuable knowledge
 514 regarding microbial diversity and community structure, underpinning the critical roles of microorganisms
 515 in the environment. However, the limitation of using DNA sequence information to infer *in situ* activity,
 516 or even potential metabolic functions, has been looming over the field of environmental microbiology from
 517 its early days. This inherent property results from both the fact that two organisms with closely-related
 518 16S rRNA gene sequences might possess different metabolic capacities (Li et al., 2019), and even if the
 519 function of the organism is known, the presence of DNA or even RNA does not necessarily indicate that
 520 the cells are active (Blazewicz et al., 2013). Recent studies are beginning to combine other types of data
 521 with amplicon sequencing to improve investigations of ecological patterns.

522 Using stable isotopes as an indicator of activity is one of the more popular and robust ways to bridge
 523 the gap between microorganisms and their function in ecological processes. In environmental microbiol-
 524 ogy, DNA or RNA stable isotope probing (SIP) is applied by incubating a sample with a isotopically-

525 labelled substrate (including heavy and rare stable isotopes of C, N, H or O), that can be incorporated
526 into the biomass of metabolically active cells (Angel, 2019; Dumont and Murrell, 2005). Unfortunately,
527 for P no stable isotopes next to the one and only ^{31}P exist. The identity/community profile of the
528 labelled organisms may then be determined using separation of different buoyant densities of the nucleic
529 acids and subsequent sequencing of the different density fractions which allows drawing causal ecological
530 interpretations of the microorganisms active in the uptake and/or assimilation of the substrate. Or-
531 ganisms labelled through SIP may further be detected and identified on a single-cell level using other
532 methods, such as Raman microspectroscopy or NanoSIMS in combination with FISH (Musat et al., 2016;
533 Wang et al., 2016).

534 Other recent advances in linking microorganisms to functions include so-called ‘next-generation physi-
535 ology’ approaches (Hatzenpichler et al., 2020). Similar to SIP, these methods require the introduction of
536 isotopically labelled or non-canonical molecule into the sample for the detection of metabolically active
537 organisms. The use of heavy-water labelling has become a recent popular approach for universal target-
538 ing of all active organisms using either $^{18}\text{O}\text{-H}_2\text{O}$ (Aanderud and Lennon, 2011; Schwartz, 2007; Angel
539 and Conrad, 2013) or deuterium oxide (D_2O) (Li et al., 2019; Eichorst et al., 2015). The assimilation
540 of $^{18}\text{O}\text{-H}_2\text{O}$ into DNA can be used to deduce microbial growth rates (Hungate et al., 2015), whereas
541 heavy water (D_2O) can be detected in the newly synthesized lipids or proteins of active cells (Li et al.,
542 2019). Combined with the identification of taxa of interest through amplicon sequencing, next-generation
543 physiology approaches represent powerful tools to bring us to the next step in soil ecological research.

544 Amplicon sequencing may also be combined with with BioOrthogonal Non-Canonical Amino acid Tagging
545 (BONCAT) to target only the fraction of cells within a soil sample that is translationally active *in*
546 *situ* (Couradeau et al., 2019; Reichart et al., 2020). The use of modified indicator molecules opens
547 new avenues for detecting metabolically active cells in the context of environmental samples, however,
548 the application to soil remains limited to very few studies so far (Couradeau et al., 2019; Reichart
549 et al., 2020). Coupling these labelling approaches to cell sorting via fluorescence-activated cell sorting
550 (FACS) (Couradeau et al., 2019) or Raman-activated cell sorting (RACS) (Lee et al., 2019), provides
551 a non-destructive alternative to NanoSIMS for identifying the metabolically active organisms, and thus
552 allowing the labelled fraction of cells to be targeted for downstream sequencing. Additionally, combining
553 these labelling approaches with cell sorting and sequencing may further circumvent challenges associated
554 with exogenous DNA.

555 In addition, amplicon sequencing can certainly also be a valuable tool for planning of more targeted
556 metagenomic or metatranscriptomic studies to investigate phylogenetic composition, functional poten-
557 tial and/or gene expression in the community context (Regalado et al., 2020). These approaches remain
558 promising for improving the link between organisms and their ecological roles and circumvent method-
559 ological challenges introduced through amplicon sequencing, such as PCR bias. However, both sequencing
560 and bioinformatic costs for gaining functionally relevant insights into ecosystem processes by “omics” ap-
561 proaches are typically orders of magnitudes higher than those needed for analyzing amplicon sequencing
562 data. The use of a limited number of metagenomes or metatranscriptomes in complement to amplicon
563 sequencing presents a cost-effective and informative approach for linking microbial community structure
564 to function in the complex soil environment.

565 8. Summary and outlook

566 Amplicon sequencing is and will remain a valuable approach for investigating the structure of microbial
567 communities in soils. However, the complex nature of soils and high diversity of organisms therein
568 necessitate careful considerations, from sampling strategies to statistical analyses, to avoid mis- or over-
569 interpretation of the data. Amplicon sequencing as a standalone approach should primarily serve as a
570 hypothesis-generation tool that is highly descriptive in nature, mainly allowing one to catalogue nucleic
571 acids of organisms present in a given sample. As one key goal of soil microbial ecology is to link organisms
572 to environmental processes, sequencing-based studies need to be complemented with other data types, in
573 addition to appropriate normalization and statistical approaches. Understanding the nature of amplicon
574 data and the role of sequencing as a valuable tool for soil scientists will further expand our understanding
575 of microbial community diversity and structure in the immensely complex soil environment.

576 Declaration of competing interests

577 The authors declare no conflict of interest.

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586 Supplementary Material

587 Hosted file

588 Supplementary_Material.docx available at [https://authorea.com/users/351324/articles/
589 475986-a-critical-perspective-on-interpreting-amplicon-sequencing-data-in-soil-
590 ecological-research](https://authorea.com/users/351324/articles/475986-a-critical-perspective-on-interpreting-amplicon-sequencing-data-in-soil-ecological-research)

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